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ST. JOHNS UNIVERSITY TO HOST THE 40th ANNUAL MACUB CONFERENCE SATURDAY, OCTOBER 20, 2007 PAUL FISHER AND SAM RHINE TO PRESENT KEYNOTE ADDRESSES



Paul Fisher

St. Johns University will host the 40th Annual Fall MACUB Conference on Saturday, October 20, 2007. The conference will feature keynote addresses by Paul B. Fisher and Sam Rhine.

Dr. Fisher is a Professor of Clinical Pathology and a Chernow Research Scientist in the Departments of Pathology and Oncology, and Neuro-Oncology Program Director at Columbia-Presbyterian Medical Center. Dr. Fisher recently discovered six novel genes that could change human melanoma cells in culture, and other human tumor cell lines, into cells that lose cancerous properties. His other undertakings include making monoclonal antibodies (MAb) specific for cancer antigens through surface epitope masking; upregulating tumor surface antigens for improved immune recognition and enhanced MAb-tumor targeting through targeted antigen modulation; and developing better approaches to identify, isolate, and clone new cancer genes, including tumor suppressors and genes involved in making cancer cells more invasive and metastatic.

Sam Rhine has attended Indiana University, Indiana School of Medicine, and Harvard Medical School. He has devoted himself to genetics education for 20+ years. He is a gifted speaker with a passion to teach the world the applicability of genetics



Sam Rhine

to daily living. Sam specializes in the most current information in such fields as Human Genetics, Biotechnology, Cloning, Stem Cells, Reproductive Biology, Prevention of Birth Defects and the Prevention of AIDS. He will present the latest technologies and information available on stem cell research.

For the past several years, Sam has concentrated on presenting Genetic Update Conferences. These one-day conferences for biology teachers and students are designed to teach the latest in genetic advances, hot research areas, and career opportunities. Sam takes biology from the textbook to the heart by posing ethical dilemmas we each will face as technological advances continue.

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Articles can be submitted electronically to invivo@mec.cuny.edu or mailed as a printed copy (preferably with a diskette that contains the file) to the Editorial Board at Medgar Evers College. All submissions should be formatted double spaced with 1 inch margins. The title of the article, the full names of each author, their academic affiliations and addresses, and the name of the person to whom correspondence should be sent must be given. As a rule, full length articles should include a brief abstract and be divided into the following sections: introduction, materials and methods, results, discussion, acknowledgments and references. Reviews and short communications can be arranged differently. References should be identified in the text by using numerical superscripts in consecutive order. In the reference section, references should be arranged in the order that they appeared in the text using the following format: last name, initials., year of publication. title of article, journal volume number: page numbers. (eg. - ¹Hassan, M. and V. Herbert, 2000. Colon Cancer. *In Vivo* **32**: 3 - 8). For books the order should be last name, initial, year of publication, title of book in italics, publisher and city, and page number referred to. (eg. - Prosser, C.L., 1973. *Comparative Animal Physiology*, Saunders Co., Philadelphia, p 59.). Abbreviations and technical jargon should be avoided. Tables and figures should be submitted on separate pages with the desired locations in the text indicated in the margins.

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The MACUB web site is now up and running. We now call for members to use the web site for registration information. Register for the 40th Annual Fall Conference on-line. Submit your poster presentation abstract on-line. Submit your member paper presentation on-line. If you are a MACUB member in good standing and have a web site that you would like linked to our web site, submit the URL address to: gsarinsky@kbcc.cuny.edu.

Color Vision and Oxygen Uptake in *Fundulus heteroclitus*

by

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Introduction

Most fishes are thought to have color vision¹. Few studies have been performed that examine the effect of background color of holding tanks on the respiration of fishes². In one study the behavior of Nile tilapia, *Oreochromis niloticus*, was examined when exposed to black, blue, green, yellow, red, and white colors. Here the highest respiratory frequency of fish occurred in a white background. This observation was made by counting opercular movements. In a study³, employing the same six colors, the VO₂ (mls O₂/gram of fish/per hour) uptake of white perch, *Morone americana*, denied access to surface waters was determined. The lowest VO₂ occurred in black vessels, with fish in green vessels having the highest VO₂.

The purpose of this study was to determine the VO₂ in the killifish, *Fundulus heteroclitus*, exposed to the colors previously mentioned.

Materials and Methods

Fish were obtained from North Pond, a fresh water pond located on Sandy Hook, New Jersey, during August and September, 2006. Fish were maintained in a university laboratory for two weeks prior to testing, in green colored tanks, at 22°C in aged tap water. Twenty-four tests, each of which included the six colors, were performed using 144 fish. Males and females were tested separately, using a single fish in each vessel. Fish were weighed (Ohaus model B300) at the conclusion of each test. The test vessels used, dissolved oxygen readings, and VO₂ determinations were made as described elsewhere³.

To determine if background color resulted in differences in VO₂, analysis of variance (alpha = 0.05) was performed for males and for females. If significant differences occurred Tukey's test of the means was performed. The same sequence of analyses were performed for males and for females, for their weights. Illumination above the test vessels and beneath the acetate cover within the test vessels was determined with a foot candle

light meter (Ex Tech Inst. #401027, Waltham Mass).

Results

The VO₂ for both males and female fish in black vessels were significantly lower (alpha = 0.05) than for those fish in red vessels. The VO₂ was lower for both sexes in black vessels than for fish in blue, green, white and yellow vessels, but not significantly different (Table 1). There was no significant difference in the weight of fish tested in the different colored vessels for either males or females (alpha = 0.05) (Table 2).

For the fishes tested, males weighed approximately 20 per cent less than the females, but had a 13 percent higher VO₂ than females (Tables 2, 3). Foot candles (FC) above the vessel covers was 25. Within the vessels FC ranged from 0.7 to 4.3 (Table 3).

Discussion

Fish in black containers of either sex had lower VO₂s than fish exposed in other colors. There was no apparent relationship to the FC. Red had the second lowest FC after black. In a study³ fish exposed in black vessels also had the lowest VO₂. However, the highest VO₂s in this study occurred in red vessels whereas the highest VO₂s in the other study occurred in green vessels³. The other study utilized white perch. This suggests that different cone arrangements are operating, and that color reception varies in different species. For example, the goldfish retina contains rods and four cone types in juveniles, and three cone types in adults⁴.

In another study using the same colors, but with a different type of apparatus it was observed that respiratory frequency was lowest in fish in a red background and significantly higher in fish in a white background². In that study opercular movements were counted to determine respiratory activity. That method is not however a quantitative measurement such as the examination of actual O₂ uptake employed in this study since the depth of respiration may result in greater O₂ uptake. For example, in one study³, reference is made to a yawning movement that occurs three to four times

Table 1. VO₂/gram/hour for male, and for female *Fundulus heteroclitus* in different colored vessels.

	Black	Blue	Red	Green	White	Yellow
VO ₂ , Males	0.18	0.22	0.24	0.22	0.21	0.20
VO ₂ , Females	0.16	0.18	0.22	0.18	0.19	0.19

Table 2. Weight (grams) comparisons for male, and for female *Fundulus heteroclitus* in different colored vessels.

	Black	Blue	Red	Green	White	Yellow
Weight, Males	5.28	4.68	4.64	4.65	4.57	4.52
Weight, Females	5.95	6.45	4.64	6.18	5.66	5.06

Table 3. Foot candles in the test vessels beneath the transparent covers.

	Black	Blue	Green	Yellow	Red	White
Ft Candles	0.7	3.7	1.5	5.1	1.1	4.3

each hour. That movement was observed for fish in all colors except for those in a black background. This movement affects the respiration rate and its resulting VO₂.

The smaller (average weight) males, except in red vessels where the average weight of males and females were the same, had higher VO₂s than females. When small and large adults of a species are compared, the total metabolism of the larger animals is higher, but the metabolic rate of the small exceeds that of the large⁵.

There was no correlation between foot candles within the vessels and VO₂ for either males or females.

Further tests of small, shallow water fish species are planned. Of interest is if the VO₂ in fish exposed to black backgrounds is consistently lower in comparison to fish exposed in other colors, and if colors other than red and green result in highest VO₂s, using this system of measurement.

Cone ratios and types, and their arrangement in the retina vary in species. Even though fishes may live in the same general habitat (e.g., shallow water) they can have different absorbance values, depending on microhabitat, niche, and other factors⁶. These factors probably affect visual responses and concomitant physiological responses, including VO₂.

References

- ¹Barton, M., 2007. Bond's Biology of Fishes, 3rd ed. Thomson, Brooks/Cole. U.S.: p. 891.
- ²Fanta, E., 1995. Influence of background color on the behavior of the *Oreochromis niloticus* (Cichlidae). *Arquivos de Biologia e Tecnologia*, **38**:1237-1251.
- ³Dorfman, D., 2007. The effect of tank color on VO₂ uptake in white perch. *In Vivo*. **28(3)**: 21-24.
- ⁴Roberts, C.M. and M.S. Loop., 2004. Goldfish color sensitivity is high under light-adapted conditions. *J. Comp. Physiol. A* **190**:993-999.
- ⁵Prosser, C.L., 1973. *Comparative Animal Physiology*, 3rd ed. W.B. Saunders Co., Pa: p. 966 and Appendix.
- ⁶Levine, J.S. and E.F. MacNichol, Jr., 1979. Visual pigments in teleost fishes: Effects of habitat, microhabitat, and behavior on visual system evolution. *Sens. Process.* **3(2)**:95-131.

A Remnant Stand of *Pinus taeda* L. the Bigwood Stand, Hertford County, North Carolina

by

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ABSTRACT

A forest containing over mature loblolly pine, *Pinus taeda* L., the Bigwood Stand, Hertford County, North Carolina, was sampled with twenty 10x 10 meter quadrats. All trees with a dbh of 7.6 cm or greater within these quadrats were sampled. *Liriodendron tulipifera* ranks first in relative dominance with a RDo of 28.3 at this site. *Pinus taeda* ranks second in relative dominance (RDo value of 21.3). The average age of the surviving loblolly pine at the Bigwood Stand was 175 years. Though *Pinus taeda* is declining, it is still a conspicuous component of the Bigwood Stand, Hertford County, North Carolina. The loblolly pine sere may persist for a longer period of time on lowland sites than predicted in current ecological texts.

Introduction

Students of ecology have long appreciated the concept of succession, an orderly process of community development that is generally directional and therefore predictable^{1,2,3}. Secondary succession occurs on sites of previous plant habitation³. Secondary succession also occurs when natural vegetation is disturbed or destroyed by fire, lumbering, windthrow or logging¹.

Billings^{4,5} studied secondary succession on the Piedmont of North Carolina near Durham, North Carolina. His dissertation at Duke University⁴ was published in Ecological Monographs⁵. Billings noted that *Digitaria sanguinalis* (crab grass) was the first annual to dominate abandoned fields. By late summer after the first year of abandonment, seeds of *Conyza canadensis* (Horseweed) were carried to the field; these germinated and by early winter produced conspicuous rosettes. Horseweed was abundant in early summer in two year abandoned fields. *Ambrosia artemisiifolia* (ragweed) was also common and was joined in late summer by perennial aster, *Aster ericoides* (white aster). *Andropogon virginicus* (broom sedge) was the dominant species during the third growing season. Pines, especially *Pinus echinata* (short leaf pine) may get established in Piedmont fields following three years of abandonment, and by ten years was the most conspicuous member of the old-field community with a density of over 7,500 individuals per hectare. Hardwoods, especially *Quercus spp.* (Oaks) become more numerous than pines by the 50th year but may not become dominant until the developing community is 100 years or older.

Many ecology texts, namely Oosting⁶ Odum² and Smith and Smith³ note the decline of pine in aging communities. Smith and Smith³ recorded that *P.*

echinata declined to near zero individuals after 110 years in old fields of the North Carolina Piedmont. Yet there are several papers that indicate that the pine sere persists for a far longer time^{7,8,9}.

Pinus taeda L., loblolly pine, is the most commercially valuable pine in the southeastern United States. *Pinus taeda* is also known as oldfield pine, North Carolina pine and Arkansas pine. Loblolly pine's natural range is from southern New Jersey south to central Florida and from Florida, west to eastern Texas. It is the most commercially important species in the coastal plain and Piedmont of North Carolina and comprises over half of the standing pine volume in the southern United States⁸.

Pinus taeda is "medium lived" pine⁸. Trees older than 200 years are rare though several investigators Stalter⁷ Baker and Landon⁸, and Pederson *et al.*⁹ have reported individual trees over 200 yrs. old. Stalter⁷ reported a giant Loblolly in the Congaree Swamp, South Carolina, with a dbh of 5' to be approximately 250 yrs. old in 1968. This tree survived Hurricane Hugo, September, 1989, and is still alive as this paper goes to press. Baker and Langdon⁸ reported a 245 yr. Loblolly pine in North Carolina, one of 20 individual trees in a stand of *P. taeda* with an average age of 240 yrs.

Hurricane and hurricane spawned tornadoes are responsible for extensive damage in southeastern forests. Hurricane Hugo, one of the most powerful storms to strike South Carolina during the 20th century, ravaged the Congaree Swamp with hurricane force wind gusts and tornadoes September 21 and 22, 1989 destroying many of the mature *P. taeda*.

Students of ecology have long-noted the role of pine in community development in the southeastern United States^{1,2,4,6,10}. The pines are gradually replaced by hardwoods especially oak (*Quercus spp.*) and

hickory (*Carya spp.*). Many ecology texts^{2,6} include this information in chapters on “community ecology/community development/plant succession.” According to Oosting⁶, a climax forest of oak-hickory is attained on abandoned farmland in the Piedmont of North Carolina after 150 yrs.

There are exceptions to the time it takes hardwoods to assume dominance in burned, abandoned, wind-throw or cut-over land in the Piedmont of the Carolinas. Stalter⁷ described a mature stand of *Pinus taeda* in the Congaree Swamp, South Carolina where most of the pine were 100-175 yrs old. *Pinus taeda* was the dominant tree in the portion of the Congaree Swamp sampled by Stalter⁷ with a relative dominance (percent basal area) value of 48.

Jones et al¹¹ described an old-growth forest stand of the Boiling Springs National Area, South Carolina. Diameter of the largest and oldest *P. taeda* here ranged from 90 to 110 cm. Two 107 cm *P. taeda* were cored and found to be 123 and 150 years old respectively. Jones et al¹¹ reported that 150 yr. old Loblolly Pine approached the maximum age attained for Loblolly Pine in the Piedmont of South Carolina. Jones et al¹¹ cited Batson et al's 1955 study at Boiling Spring¹² where there were score or more *P. taeda* with trunk diameters ranging to an excess of 1.5m. Twenty five years later, in 1980, the number of large *Pinus taeda* at Boiling Spring had been reduced to ten individuals; over half of the large *P. taeda* reported by Batson et al¹² died.

Pederson et al.⁹ discussed the age structure and possible origins of old *Pinus taeda* stands in the Congaree Swamp, South Carolina. Two of the stands in the Pederson et al⁹ study, stand six and seven, contained large old *P. taeda*. The average age of the *P. taeda* cored in stand six was 137 years with a range of 122-156 years while average age in the stand seven was 157.5 yrs, with a range of 124-227 yrs.

Pederson et al.⁹ reported the there were two or more age classes of *Pinus taeda* in the Congaree Swamp, South Carolina. Pederson et al⁹ claimed that Swails et al¹³ and Stalter⁷ reported even aged stands of *Pinus taeda* in the Congaree Swamp. Yet the second line of the abstract of Stalter's⁷ Congaree Swamp paper states, “Annual ring counts indicated that most trees (*Pinus taeda*) are 100-175 years old.” Stalter⁷ also reported in the first sentence, second paragraph, of the results section, “Most of the *P. taeda* in the stand are from 60 to 98 cm in diameter. Fifty borings

and the annual ring counts on the stumps of three cut pine indicate that most individual *P. taeda* are 100-175 years old; a few trees are older.” In spite of the aforementioned information, Pederson et al⁹ claimed that Stalter⁷ described, “an even aged stand,” of *Pinus taeda* in the Congaree Swamp.

Methods

Trees in the Bigwood Stand with a dbh 7.6 cm or greater were sampled in twenty 10x 10m quadrants (Table 1). An additional stand of *Pinus taeda*, a 110 yr stand located within Duke Forest, Durham, North Carolina was also sampled by the twenty 10 by 10m quadrats. Relative dominance data, RDo, for the trees found within the “Duke Forest” is presented in Table 2. Density and frequency data for *Pinus taeda* stands varying in age from 9 yrs to 180 yrs are presented in Table 3. Nomenclature follows that of Radford et al¹⁴.

Results and Discussion

Acer rubrum, the most abundant tree at the study site, had the highest importance value, 46.50, *Liriodendron tulipifera* attained the highest relative dominance (RDo) value followed by *Pinus taeda* (Table 1). Other trees listed in order of decreasing relative dominance are *Liquidamber styraciflua*, *Quercus falcata* and *Carya spp.* *Acer rubrum* ranks first in relative density followed by *Carya sp.*, *Quercus falcata*, *Oxydendron arboreum* and *Liquidamber styraciflua*. *Pinus taeda* ranks 8th in relative density.

Living *P. taeda* were observed in four quadrats (Table 3). Dead loblolly pine, either standing or fallen were observed in six quadrats, including five quadrats where no living Loblolly pine were observed. One dead standing *P. taeda* with dbh of 117cm was the largest loblolly pine in North Carolina when it was alive.

Most of the living *P. taeda* in the stand ranged from 35 to 93cm in diameter. Annual ring counts provided by the International Paper Co. on living *P. taeda* in 1995 indicated that most *P. taeda* were over 175 years old. Although the living loblolly pine appeared to be healthy, there were more dead standing or recently fallen *P. taeda* at the site than living trees.

Data presented in Table 2 shows the relative importance of *P. taeda* to hardwoods in old field succession on lowland sites. The data for stands 15 to 90 years old were transposed from counts made by Oosting¹. Data for the 110 yrs stand are from the Mud Creek site, Duke Forest,

Table 1. Density, (D) Relative Density (RD) Frequency (F) Relative Frequency (RF) Basal Area (BA) Relative Dominance (RDo) and Importance Value for trees in the Bigwood Stand, Hertford County, North Carolina.

Species	D	RD	F	RF	BA	RDo	IV
<i>Acer rubrum</i>	1.05	21.8	65	21	358.96	3.7	46.5
<i>Liriodendron tulipifera</i>	0.3	6.3	25	8.1	2746.53	28.4	42.8
<i>Liquidambar styraciflua</i>	0.5	10.4	30	9.7	1832.34	18.9	39
<i>Quercus falcata</i>	0.6	12	40	12.9	1302.98	13.5	38.9
<i>Pinus taeda</i>	0.2	12.5	20	6.5	2058.53	21.3	32
<i>Carya spp.</i>	0.7	4.2	32	11.3	560.79	5.8	31.7
<i>Oxydendron arboreum</i>	0.6	1406	45	14.5	342.42	3.5	30.5
<i>Quercus alba</i>	0.45	12.5	20	6.5	111.52	1.2	17.1
<i>Cornus florida</i>	0.1	9.4	5	1.6	14.14	0.1	3.8
<i>Carpinus caroliniana</i>	0.1	2.1	5	1.6	62.84	0.7	4.4
<i>Ilex opaca</i>	0.005	2.1	5	1.6	98.18	1	3.6
<i>Fraxinus spp.</i>	0.005	1	5	1.6	63.62	0.7	3.3
<i>Quercus michauxii</i>	0.005	1	5	1.6	50.27	0.5	3.1
<i>Quercus phellos</i>	0.005	1	5	1.6	78.54	0.8	3.2
			99.9	100.1	9881.66	100.1	

Table 2: The relative importance of *Pinus taeda* and hardwoods in old field succession on lowland sites.

Age of Stand (Years)	Relative Density (%)	Relative Dominance (%)
15	92	99
18	89	99
34	30	88
45	22	64
90	15	76
110	23	76
180	4.2	21.3

The data for all stands except the 110 and 180 year old stands were transposed from the counts made by Oosting (1942).

Table 3. Density and Frequency of *Pinus taeda* in a successional series of seven bottom land pine stands.

Age of Stand (Years)	Density	Frequency
15	20.3	100
18	19.3	100
34	7.9	100
45	4.9	100
90	3.0	90
110	1.1	95
180	0.2	20

Durham North Carolina measured by Stalter and Dial in 1997. Pine comprised almost all the percent basal area (RDo) in 15 and 18 yrs stands, and approximately 75 percent of the relative dominance of stands of 100 yrs old. Although the density and frequency of *P. taeda* declines from 15 to 110 yrs, relative dominance declined very slowly (Tables 2 and 3). At sites over 150 yrs, *P. taeda* has low density and frequency values, (Table 3) but because of its large trunk diameter (d_{hb}) *P. taeda* is still an important component of the forest (Table 2).

Pomeroy (1968 personal communication) stated that *P. taeda* in the Bigwood Stand, Hertford County, North Carolina averaged 145 yrs when cored in 1950. While the Bigwood Stand was thriving at the sites in 1968⁷, *P. taeda* has declined rapidly in the past 30 years⁷. There were more dead standing trees, and fallen *P. taeda* present at the site in 2002 than living Loblolly Pine. The data in this study suggest that *Pinus taeda* may remain dominant on lowland sites in the southern United States, for 150 years or longer if it can escape logging, wind throw, insect infestation and if it can escape cutting^{7,9,12,15}.

References

¹Oosting, H.J., 1942. An ecological analysis of the plant communities of Piedmont, North Carolina. Amer. Midland Natur. **28**: 1-126.
²Odum, E.P., 1971. Fundamentals of Ecology. 3rd ed. W. B. Saunders Company 547 p.
³Smith, T.M. and R.L. Smith, 2006. Elements of Ecology. 6th ed. Pearson Ed., Inc. publishing as Benjamin Cummings. San Francisco. 658 p.
⁴Billings, W.D., 1936. The structure and development of old field short-leaf pine stands and certain associated physical properties of the soil. Ph.D. Dissertation, Duke University, Durham, N. C. 148 p.

⁵Billings, W.D., 1938. The structure and development of old field short-leaf pine stands and certain associated physical properties of the soil. Ecological Monographs **8**: 437-499.
⁶Oosting, H.J., 1956. The Study of Plant Communities. 2nd ed. Freeman and Co., San Francisco, Calif. 440 p.
⁷Stalter, R., 1971. Age of a mature (*Pinus taeda*) stand in South Carolina. Ecology **52**: 532-533.
⁸Baker, J.B. and O.G. Langdon, 1990. *Pinus taeda* L. Loblolly Pine. Silvics Manual of North America Volume 1. Conifers. Forest Service, United States Department of Agriculture, Washington, DC 497-512.
⁹Pederson, N.A., R.H. Jones and R.R. Sharitz, 1997. Age structure and possible origin of old *Pinus taeda* stand in floodplain forests. Journal Torr. Bot. Society **124**: 111-123.
¹⁰Quaterman, E. and C. Keever, 1962. Southern mixed forest: Climax in the southeastern Coastal Plain, USA. Ecol. Mongor. **32**: 167-185.
¹¹Jones, S.M., D.H. Van Lear and S.K. Cox, 1981. Competition and diameter pattern of an old growth forest stand of the Boiling Springs Natural Area, South Carolina. Bull. Torrey Bot. Club **180**: 347-353.
¹²Batson, W.T., W.R. Kelley, L.F. Swails and F.F. Welbourne, 1955. The vegetation of a mature beech-magnolia forest within the Gant Tract. University of South Carolina, Publ. services III 65-71.
¹³Swails, L.F., W.D. Anderson Jr. and W.T. Batson, 1957. A mature pine stand in the Congaree bottom land. University of South Carolina Publ. Biol. Ser. III: 82-89.
¹⁴Radford, A.E., H.E. Ahles and C.R. Bell, 1968. Manual of the Vascular Flora of the Carolinas. University of North Carolina Press. Chapel Hill.
¹⁵Keever, C., 1950. Causes of succession in old fields of the Piedmont, North Carolina. Ecol. Monogr. **20**: 97-105.

The Effect of Grape Seed Extract on Hematopoietic Stem Cells in the Umbilical Cord Blood

by

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ABSTRACT

The human umbilical cord blood is known to possess more progenitor and hematopoietic cells than peripheral blood and is a very good candidate for studying the effect of natural health substances (NHSs) on mononucleocytes (MNC) subsets. Cells capable of initiating the human cell engraftment (Severe-Combined-Immunodeficiency-Repopulating cells) are contained in the CD34⁺ cell fraction, which was the focus of the current study in the light of the effect of grape seed extract on hematopoietic stem cells in the human umbilical cord blood. The target cells were from the human erythroleukemia cell line K562 cells. Cell line K562 CLL 220 is the model karyotype of the human leukemia cell line, K562 (ATCC CCL 243).

Introduction

A well planned systemic approach for treating patients with cancer should aim at the multiple physiological and biochemical pathways, which involve tumor evolution. Such an approach should also control and restrict any damages to normal tissues. Angiogenesis is a basic course in the enhancement of cancer cells. Various natural health substances are reported not only to decrease angiogenesis but also described to have the potential of other anticancer properties. Herbalist have provided hundreds of years of remarkable information on the anticancer pursuit of numerous natural health substances. Natural health substances (NHSs), such as Resveratrol and Proanthocyanidin (grape seed extract, GSPE), not only are reported to have a high

degree of antiangiogenic activity but also other numerous relationships that can hinder tumor series and decrease the risk of metastasis. NHSs target various molecular pathways other than angiogenesis, including vascular endothelial growth factor (VEGF)¹, epidermal growth factor receptor (EGFR), the HER-2/neu gene, the protein kinases, coagulation pathways and others. Laboratory studies are as well strengthening these relationships, which are already recorded in the customary texts. The following NHSs are routinely used for anticancer therapy and are antiangiogenic for the sake of multiple co-dependent courses: Quercetin dihydrate, Astaxanthin, Ginsenoside Compound-K (CK-1), *Cordyceps Sinensis* (TCS-W), Caffeic acid, Bilobalide, Caffeic acid, Eugenol, Rutin hydrate, wheat grass extracts, resveratrol, GSPE, and *Ganoderma lucidum*^{2,3}. Grape

seed extract containing GSPE and other antioxidants are being used as nutritional supplements by many health conscious individuals. The advantageous effects of GSPE have been reported, however, still little is known about their mechanism(s) of action. One of the beneficial effects of GSPE is chemoprevention of cellular damage. The exact mechanism by which GSPE mediates chemoprevention is not yet understood. Exposure to GSPE results in a significant reduction in apoptosis⁴. Conversely, VEGF is believed to be the most prevalent, efficacious, and long-term signal that is known to stimulate angiogenesis in wounds. The latter is rich in oxidants such as hydrogen peroxide contributed by neutrophils and macrophages. GSPE facilitates oxidant-induced VEGF-expression in keratinocytes. In fact, GSPE is a group of biologically active polyphenolic bioflavonoids that are synthesized by many plants. GSPE contain 5,000 ppm Resveratrol. GSPE can as well up-regulate hydrogen peroxide¹.

We investigated the phenomena of hematopoietic stem cells in the human umbilical cord blood cells, which were pretreated with the extract of GSPE. Up to 78% to 100% of the human umbilical cord blood cells express the hematopoietic progenitor-cell-surface- marker CD34, which can consistently engraft, develop, and proliferate in the hematopoietic tissues of sublethally irradiated severe-combined-immunodeficiency (SCID) mice^{5,6,7,13}. Data suggest that cells capable of initiating the human cell engraftment, namely, SCID-repopulating cells are contained in the CD34⁺ cell fraction, which was the focus of the current study. CD34 not only can be found on the hematopoietic stem cells; while conversely, CD45 can be identified on all nucleated hematopoietic cells.

Materials and Methods

This study was approved by the Internal Review Board of the Hospital, where babies were delivered. The study as well conformed to the Declaration of Helsinki. Each of the six individuals who provided umbilical cord blood gave written consent. The umbilical cord blood

(UCB) was taken under aseptic technique from the umbilical vein after the full term baby was delivered and before the placenta was separated from the uterus.

Technique of Magnetic-bead Cell Separation

CD34CD45⁺hematopoietic stem cells separated from the UCB mononucleocytes (MNCs) were augmented with the procedure of a positive magnetic-bead cell separation (MACS, *Miltenyi Biotec, US*). We adapted the technique of Gritzapis, *et al.*¹¹. The magnetically tagged cells were placed into two MACSy RS1 separation columns (Miltenyi Biotec), which had been equilibrated with the buffer solution in the magnetic field of the Vario MACSy separator (Miltenyi Biotec, US). The columns were then prepared by rinsing with buffer solution. Retained cells were extracted from the column to ascertain that they were away from the magnetic field. Aliquots of the selected cells were then stained by the PC5/FITC labelled anti-CD34CD45⁺ monoclonal antibodies (Coulter Immunotech, US) to examine the purity of CD34CD45⁺ stem cells. This purity was ascertained by flow cytometry analysis.

After enrichment of CD34⁺ hematopoietic stem cells by FACS BD (Fluorescence Activated Cell Sorting) to isolate/identify cells expressing with a certain surface marker, the hematopoietic stem cells were incubated for 7 days with six different concentrations of GSPE ranging from 10 to 100 µg/mL. The size of effector cells was assessed using Centricon (cut-off: 30 kDa, Amicon Millipore Co. US).

The Preparation of Monoclonal Antibodies

All monoclonal antibodies to surface antigens such as CD34, CD 45, CD26, CD19, CD14, and CD3 (FITC; Serotec, US) were obtained from Coulter Immunotech, US. We chose the following for this study: CD34⁺-Enriched Cells originated from the umbilical cord blood⁵, CD45⁺ lymphocytes and granulocytes, CD14-a monocyte/macrophage differentiation marker⁸, and CD26- a cell surface protease, expressed on many cells of

the immune system including some CD4⁺ T-cells and macrophages⁹.

The Activation of Effector Cells

Highly augmented CD34CD45⁺ stem cell suspensions were cultured in a medium supplemented with RPMI-1640 for 24 hours (37°C, 10% CO₂). Six different concentrations of GSPE and of positive controls respectively, ranging from 10 µg/mL to 100 µg/mL by serial dilution, were added to cell suspensions for preincubation treatment prior to the subsequent cytotoxicity test. The control group was treated with phosphate buffered saline (PBS). After 7 days of incubation, the cultures were washed with PBS and then re-suspended in a medium containing 20% fetal bovine serum (FBS), at which point they were ready for the cytotoxicity assay.

The Expression of Stem-Cell Mediated Cytotoxicity

The results of the effect of GSPE on hematopoietic stem cell mediated cytotoxicity were expressed as ratios of survival of K562 cells of GSPE treated groups versus the controls (K562 cells).

The Preparation of Target Cells

We used the human erythroleukemia cell line K562 (CCL-243, American Type Culture Collection (ATCC) as an MNCs-sensitive target for cytotoxicity assays. The percentage of lyses against target cells was measured and reported as the ratio of sample % vs. control % of cell gated in flow cytometry. Scores of 1.50 or higher were considered significant.

To detect target cell survival, we used Alamar Blue, a colorimetric indicator, which after uptake by living cells changes from a oxidized, non-fluorescent, blue state to a reduced, fluorescent, red state.

Flow Cytometric Attainment

Flow cytometry was performed with a FACScalibur cytometer (Becton Dickinson). The instrument was set for two-colour analysis

using FACScomp software and was calibrated using Calibrite beads (Becton Dickinson) with a threshold of 200 on Forward Scatter (FSC) to exclude debris. Data were collected in listed mode and analyses were performed using CellQuest software version 3.1f (Becton Dickinson) and Win MIDI version 2.8 software. At least 10,000 target cells were collected and analyzed. During the procedure, CD34, CD19, CD14, and CD3 were used against lymphocyte, and their surfaces were marked. Fluorescent markers, PC5-labeled methods and Fluorescein 5-isothiocyanate (FITC) labeled methods were used (Figs. 1, 3). For the cytotoxicity analysis, the tests were done at the different effector/target cells ratios of 5:1, 20:1 and 80:1. A two tailed p value < 0.005 indicates a statistically significant difference.

Results

After treating the mononuclear cells (MNCs) of the human umbilical cord blood (hUCB) with crude extract of NHSs, we observed the ratio change of sample % vs. control % gated in flow cytometry. Flow cytometric analysis of four different CDs (CD 34, CD 19, CD14, and CD 3) revealed that among CD34 hematopoietic stem cells, the ratio of samples % vs. controls % after the treatment with GSPE was 2.18, which was the highest among all the eleven NHSs. By comparison, the ratio after treatment with the extract of wheat grass was 0.61, which was the lowest, (Fig. 1).

Among the 11 NHS, there were 52 measures having the ratio of sample % vs. control % of cell gated in flow cytometry less than 1.50, while three greater than 1.50, (Fig. 2). The t test for independent samples was done. It revealed that the difference between these two groups was statistically significant (two tailed p < 0.005).

The purity of CD34CD45⁺ stem cells was ascertained by flow cytometry analysis and scored up to 95%. The cytotoxicity at an effector/target (E/T) ratio of 5:1 was not significant, compared to that of the controls. Conversely, when E/T ratios were as high as 80:1, no high cytotoxicity effect was observed.

Such a phenomena was likely due to the over-saturation of cell numbers (data not shown). The highest level of cytotoxicity was noted at an E/T ratio of 20:1 when the effector cells were pre-incubated with 100 µg/ml GSPE, but not those at an E/T ratio of 5:1, and 80:1, respectively.

It is noted that there was individual variance of different fractions of stem cell size. Stem cell cytotoxicity increased by 1.57 fold while the negative control was 2.13 ($p < 0.05$) for the fraction of cells sized > 30 kDa, whereas 1.07 fold ($p < 0.05$) for those sized < 10 kDa after pre-treatment with 100 µg/ml of the GSPE extract. Note that, when compared with the untreated control, the negative control was 2.13. This ratio was significant for the former, > 30 KDa fraction of stem cells (1.57), while not significant for the latter, < 10 Kda fraction (1.075), according to Nociari et al's method of separation¹² (Fig. 3).

By comparison, cells after enrichment of CD14CD26⁺ MNCs/Macrophages revealed

that their cytotoxicity increased significantly by the aforementioned E/T ratio of 1.51, and 1.613 ($p < 0.01$) in pre-incubation with 1 µM and 10 µM, respectively of TPA (*a priori* tumor promoter 12-O-tetradecanoyl phorbol-13 acetate), which had served as positive controls (Data not shown).

Discussion

The human umbilical cord blood is known to possess more progenitor cells than peripheral blood, and is an excellent candidate for studying the effect of NHSs on MNC subsets. The UCB samples from six different individuals after treatments with GSPE exhibited changes in the ratio of sample % vs. control % gated in flow cytometry, predominantly for the hematopoietic stem cells with CD34CD45 staining in the current study. Recognizing the repopulating characteristics of the human hematopoietic stem/progenitor cell fractions is vital for predicting their

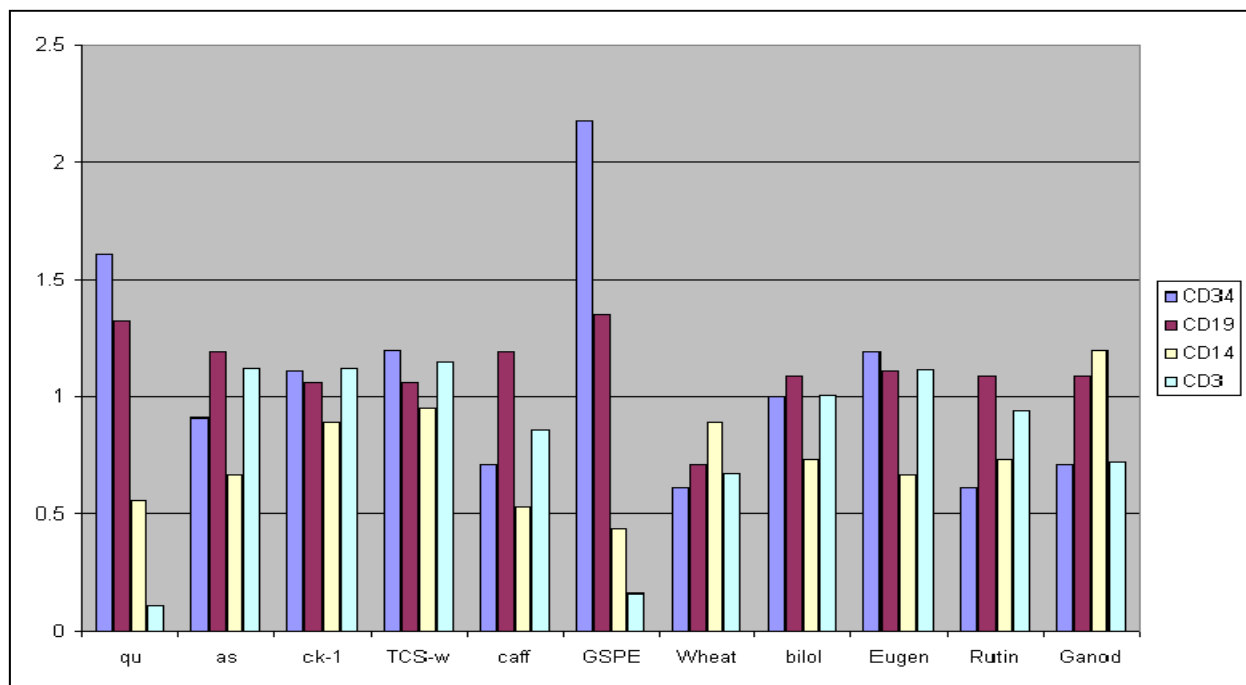


Fig. 1. Flow cytometric analysis of four different CDs (CD 34, CD 19, CD14, and CD 3) revealed that among CD34 hematopoietic stem cells, the ratio of samples % vs. controls % after the treatment with GSPE was 2.18 (See Y- Axis), which was the highest among the eleven NHSs (See X-Axis). By comparison, the ratio after treatment with the extract of wheat grass was 0.61, which was the lowest. CD34 can be found on hematopoietic stem cells; while CD45 can be identified on all nucleated hematopoietic cells.

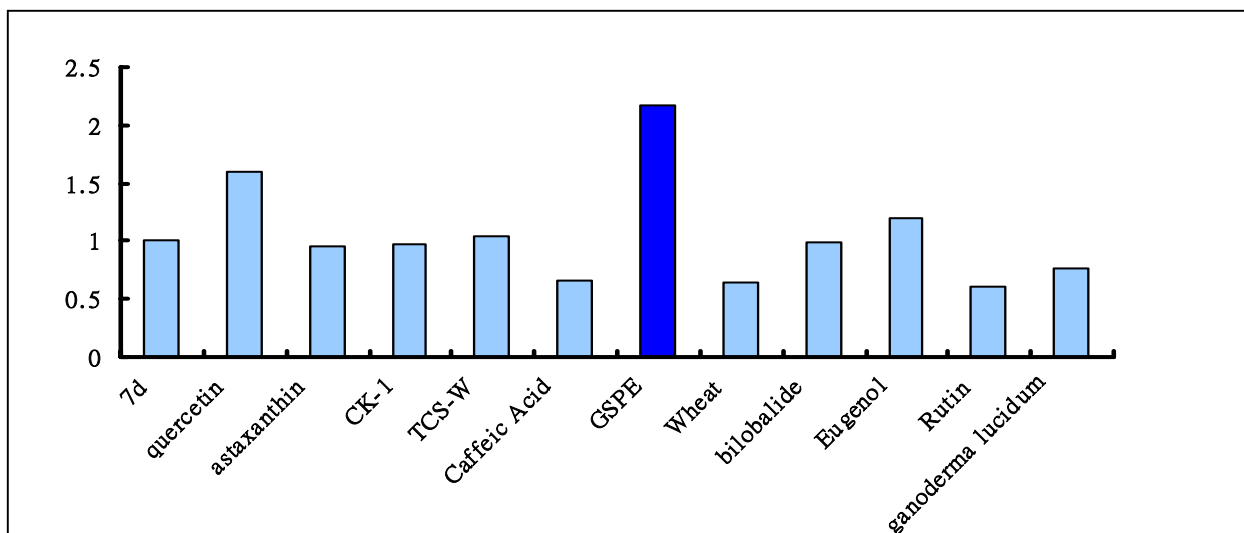


Fig. 2. Among 11 natural health substances (See X-Axis), there were 52 measures having the ratio of sample % vs. control % of cell gated in flow cytometry less than 1.50, while the remaining three greater than 1.50.

performance, for instance, after transplant into high-risk patients following high-dose of radiation therapy. Nevertheless, our preliminary investigation did have some limitation, which may temper our conclusion and should be addressed in future study. For instance, technically, the long diagonal lane of events in the flow cytometry seems to include the sheer quantity of some dead cells, which may in turn affect the counting of precisely real CD34-stained cells. Hence, some experimental aspects must be more profoundly assessed for future evaluation.

Conclusion

On balance, with new developments¹⁴, the future identification and purification of the molecule(s) of the crude extracts of GSPE, which are active for augmenting development and proliferation of the hematopoietic stem cells in the human umbilical cord blood, may open a new avenue of managing angiogenesis, and conversely, may as well potentially promote the oxidant-induced VEGF-expression.

References

- ¹Khanna S, S. Roy, D. Bagchi, M. Bagchi, and C.K. Sen, 2001. Upregulation of oxidant-induced VEGF expression in cultured keratinocytes by a grape seed proanthocyanidin extract. *Free Radic Biol Med.* **31(1)**:38-42.
- ²Yance, D.R. and S.M. Sager, 2006. Targeting angiogenesis with integrative cancer therapies. *Intergative Cancer Therapies* **5(1)**: 9-29.
- ³Chien, C.M., J.L. Cheng, W.T. Chang, M.H. Tien, C.M. Tsao, Y.H. Chang, H.Y. Chang, J.F. Hsieh, C.H. Wong and S.T. Chen, 2004. Polysaccharides of *Ganoderma lucidum* alter cell immunophenotypic expression and enhance CD56+ NK-cell cytotoxicity in cord blood. *Bioorg Med Chem.* **12**: 5603-5609.
- ⁴Joshi, S.S., N.N. Babushkina-Patz, D.J. Verbik, T.G. Gross, S.R. Tarantolo, C.A. Kuszynski, S.J. Pirruccello, M.R. Bishop and A. Kessinger, 1998. Antitumor activity of human umbilical cord blood cells: a comparative analysis with peripheral blood and bone marrow cells. *Int J Oncol* **13**:791-799

- ⁵Hogan, C.J., E.L. Shpall, O. McNulty, I. McNiece, J.E. Dick, L.D. Shultz and G. Keller, 1997. Engraftment and Development of Human CD34+-Enriched Cells From Umbilical Cord Blood in NOD/LtSz-scid/scid Mice. *Blood* **90(1)**: 85-96.
- ⁶Ohmori, T, J. Mimuro, K. Takano, S. Madoiwa, Y. Kashiwakura, A. Ishiwata, M. Niimura, K. Mitomo, T. Tabata, M. Hasegawa, K. Ozawa and Y. Sakata, 2006. Efficient expression of a transgene in platelets using simian immunodeficiency virus-based vector harboring glycoprotein I α promoter: in vivo model for platelet-targeting gene therapy. *FASEB J.* **20(9)**:1522-1524.
- ⁷Freitas, C.S and S.R. Dalmau, 2006. Multiple sources of non-embryonic multipotent stem cells: processed lipoaspirates and dermis as promising alternatives to bone-marrow-derived cell therapies. *Medicine. Cell and Tissue Research* **325(3)**:1432-0878.
- ⁸Steiner A.A., S. Chakravarty, A.Y. Rudaya, M. Herkenham, A.A. Romanovsky, 2006. Bacterial lipopolysaccharide fever is initiated via Toll-like receptor 4 on hematopoietic cells. *Blood* **107(10)**: 4000-4002.
- ⁹Morimoto C and S.F. Schlossman, 1998. The structure and function of CD26 in the T-cell immune response. *Immunol Rev.* **161**:55-70.
- ¹⁰Shizuru, J.A., R.S. Negrin and I.L. Weissman, 2005. Hematopoietic stem and progenitor cells: Clinical and Preclinical Regeneration of the Hematolymphoid System. *Annual Review of Medicine* **56**: 509-538.
- ¹¹Gritzapis, A.D., D. Dimitroulopoulos, E. Paraskevas, C.N. Baxevanis and M. Papamichail, 2002. Large-scale expansion of CD3+CD56+ lymphocytes capable of lysing autologous tumor cells with cytokine-rich supernatants. *Cancer Immunol Immunother* **51**:440 - 448..
- ¹²Nociari, M.M., A. Shalev, P. Benias and C. Russo, 1998. A novel one-step, highly sensitive fluorometric assay to evaluate cell-mediated cytotoxicity. *J Immunol Methods* **213**:157– 167. .
- ¹³Horn, P.A., H. Liem, A. Kiyoshi and T. Yahata, 2006. Expansion of severe combined immuno deficiency (SCID) does not prove expansion of hematopoietic stem cells. *Blood* **108**: 717-772.
- ¹⁴Joshi, S.S., C.A. Kuszynski and D. Bagchi, 2001. The Cellular and Molecular Basis of Health Benefits of Grape Seed. Proanthocyanidin Extract. *Current Pharmaceutical Biotechnology*, **2(2)**: 187-200.

Bioaccumulation and Tissue Distribution of Arsenic, Cadmium, Copper and Zinc in *Crassostrea virginica* Grown at Two Different Depths in Jamaica Bay, New York

by

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ABSTRACT

Historically, Jamaica Bay was a site of extensive oyster beds and shellfish culture leases that supported a significant oyster fishery in the New York area. The industrial and urban expansion of the early 1900's led to over-harvesting and a deterioration in water and bay sediment quality that coincided with shellfish decline and the ultimate disappearance of oysters from the bay. Over the past 50 years, efforts to arrest and reverse the pollution problems of Jamaica Bay have been undertaken but the area still contains metals and other pollutants at levels higher than NYS Water Quality Standards. Previous we showed that *Crassostrea virginica* seed transplanted to the bay had excellent growth and survival despite the bay's pollution problems. In this study we measured the one-year bioaccumulation and tissue distribution of four metals in *C. virginica* seed that were transplanted to the bay at two different depths: one foot from the surface and one foot above the sediment. Tissues of *C. virginica* were dissected, dried and digested in nitric acid. Arsenic, cadmium, copper and zinc levels were measured using electrothermal vaporization with deuterium lamp background correction in an atomic absorption spectrophotometer fitted with a THGA graphite furnace. Metals were distributed in the various tissues in $\mu\text{g/g}$ dry weight amounts, which correlate well with published values for whole oysters grown in other polluted areas. Metal distributions were not homogeneous throughout the animals and in most of the tissues tested, oysters grown near the surface accumulated more metal than those positioned near bay sediment.

Introduction

Jamaica Bay, a 26 square mile estuarial embayment situated between southern Brooklyn and Queens, NY and a major inlet opening to the Atlantic Ocean, lies just east of the entrance to NY Harbor and the mouth of the Hudson River. It is home to the Jamaica Bay Wildlife Refuge, which encompasses 9,155 acres of diverse habitats including salt marsh, upland field and woods, fresh, and brackish water ponds, and an open expanse of bay and islands. The bay is also a critical component of a larger watershed that drains naturally or via storm sewers, on the seaward-sloping outwash plain south of the harbor hill terminal moraine¹.

At one time, wild stocks of the Eastern Oyster, *Crassostrea virginica*, also known as

the American Oyster, were found all along the Atlantic and Gulf coasts of North America and for centuries, supported subsistence fishing by Native Americans and early European colonists². Historically, *C. virginica*, flourished in Jamaica Bay and the NY/NJ Harbor area as either self-sustaining or farmed populations³. Jamaica Bay's oyster industry observed a steady decline in production after its peak in the early 1900's⁴. Lack of adequate supply of seed oysters, over-harvesting by commercial fishermen, increased pressure from natural predators, parasitic invasion, changing hydrographic patterns, siltation, and a decline in water quality are all cited as possible causes for the decline. The rapid urbanization and local industrialization of the area at the turn of the century was followed by years of unregulated

industrial, urban and residential dumping which contaminated bay sediment, causing adverse effects on benthic organisms and bioaccumulation further up the food chain. Discharges of inadequately treated sewage were poisoning oysters, clams and ultimately people, and by 1921 the U.S. Department of Agriculture had closed shellfish lands in Jamaica Bay altogether. The shellfish problem was not unique to Jamaica Bay for around the same time declines in estuarine shellfish populations had occurred throughout the entire east coast of the United States and other important oyster fisheries, like that of Chesapeake Bay and Virginia's James River, had also started to collapse^{5,6,7}. Today, very few if any wild oysters are found in Jamaica Bay, and the dramatic loss of this historic oyster bed has permanently altered the structure and function of the bay's benthic ecosystem.

Over the last 50 years, major efforts have been undertaken by federal, state and local authorities to arrest and reverse the pollution problem and while natural stocks of *C. virginica* have not returned to the bay, there has been a resurgence of many other marine organisms. Despite improvements, Jamaica Bay still exhibits poor water quality⁸ and studies indicate the presence of various metal pollutants, including arsenic, cadmium, chromium, copper, iron, lead, mercury, nickel and zinc, in bay sediment^{9,10,11,12,13,14,15}. Heavy metals have been found to inhibit growth in a variety of mollusc species including oysters^{16,17,18}. In 2002, our lab did a study to determine the growth and survival of *C. virginica* seed transplanted to Jamaica Bay in protected containers at two ecologically different locations, at two different depths: one foot below the surface and one foot off the bottom. Despite bay conditions, the oysters at both locations and both depths had excellent growth and survival rates¹⁹. The results of this project generated many new questions such as: to what extent are oysters acquiring bay pollutants in their tissues, how are they able to adapt to this stress, what are the physiological effects on their various organ systems, and what are the

consequences of the polluted environment on the long term survival and reproductive success in Jamaica Bay.

Bivalves are particularly good accumulators of heavy metals^{20,21,22,23} and being sessile, tend to reflect local contaminant concentrations more accurately than crustaceans and free-swimming finfish. The EPA considers *C. virginica*, as one of six target bivalve species recommended for contaminant monitoring in Mid-Atlantic coastal waters including Jamaica Bay²⁴. Target metal analytes on the EPA Fish Contaminant Workgroup list include arsenic, cadmium, mercury, selenium and tributyltin²⁵. The Mussel Watch Project of the National Oceanic and Atmospheric Administration (NOAA) monitors copper, lead, nickel, and zinc, in addition to the analytes on the EPA's target list²⁶. Numerous other reports have been made on the bioaccumulation of these and other heavy metals in oyster species of in the US and around the globe^{27,28,29,30,31,32}.

In this study we measured the one year bioaccumulation and tissue distribution of four metal pollutants (arsenic, cadmium, copper and zinc) in *C. virginica* seed, that had been transplanted to the bay in 2002 at two different depths: in floats, one foot below the surface, and in hanging nets, one foot off the sediment. Since Jamaica Bay sediment is a major reservoir of metal pollutants, it was hypothesized that metal accumulations might be greater in the tissues of bottom positioned oysters, since they were grown nearer to a more concentrated source of metal pollutants.

Materials and Methods

In June 2002, four modified Taylor Floats³³ of approximately 3' x 4' were constructed using PVC tubing and 1/4" mesh nylon screening. Each float was designed to hold up to three 1/8" mesh nylon boxes in which oyster seed were to be placed. Each float had 1/4" nylon mesh lids to keep out predators. Oyster seed of approximately 20 mm antero-posterior (height) shell lengths and 5 mm shell hinge width were obtained from Frank M. Flower & Sons, Inc. Oyster

Nursery in Oyster Bay, NY. 150 oyster seeds were distributed among the 3 nylon boxes in each float. Two floats were positioned 1 foot below the water surface in Jamaica Bay at Fort Tilden at Gateway National Park Marine Station (GNPMS), and at Kingsborough Community College Marina (KBCCM) in Brooklyn's Sheepshead Bay, a large cove of Jamaica Bay (NY/NJ Baykeeper license #LG P00 583). The two other floats that were designed to sink and position the oyster seed 1 foot off the bay bottom, were placed at the same two sites. Both surface and bottom floats were inspected and cleaned of any fouling, biweekly in the summer and monthly in the winter. After 2 months, the bottom floats were deemed clumsy/difficult to work with and were replaced with commercially constructed hanging nets, suspended one foot off the bottom, for the rest of the experiment.

Seed survival and growth, as well as various parameters of bay water quality including temperature, pH, dissolved O₂, chlorophyll a, salinity, and turbidity were monitored throughout the one year experimental period. In July 2003, samplings of one-year old oysters were removed from each site and depth to determine the bioaccumulations of arsenic, cadmium, copper and zinc in various tissues.

Prior to tissue isolation, all glassware was acid-washed in dilute (5%, wt/wt) nitric acid in deionized water to remove any bound metals. Washing was followed by a thorough rinse with deionized water to remove any remaining acid. All acids used were Fisher trace-metal grade. Representative oysters were cleaned, shucked, and their tissues dissected to determine metal bioaccumulations. The mantle, gill, palps, posterior adductor muscle and stomach were removed with stainless steel instruments. The tissues were blotted and placed in pre-weighed Pyrex flasks to determine wet weights. A 1-2 g sampling of each animal's shell was also removed, weighed, and processed similarly for metal analysis. All tissues were dried in an oven at 120°C and reweighed to determine dry weights. Dried tissue samples were digested

in concentrated nitric acid on hotplates. Digested samples were adjusted to a final volume of 10 ml in dilute (0.2 %, wt/wt) nitric acid. Aliquots of each sample were analyzed for arsenic, cadmium, copper and zinc determinations by electrothermal vaporization with deuterium lamp background correction in a Perkin Elmer AAnalyst 800 Atomic Absorption Spectrophotometer with a THGA Graphite Furnace. Metal levels were recorded as µg/g dry weight. Statistical analysis was determined by a Wilcoxon Matched-Pairs Signed Ranks Test using GraphPad InStat version 3.00.

Results

C. virginica at both locations and depths were deemed to have excellent growth and survival. After one year, oyster seed at both sites and depths had over 75% survival, and growth, as measured by shell height, had increased over 300% with average growth rates for bottom maintained oysters exceeding those maintained near the surface by approximately 20%¹⁹.

Figures 1-4 show the one-year bioaccumulations of arsenic, cadmium, copper and zinc in the mantle, gill, palps, adductor muscle, stomach and shell of top positioned oysters compared to bottom positioned oysters. For each metal, top oysters from both sites (GNPMS and KBCCM) were combined, and bottom oysters from both sites were combined, to generate top/bottom comparisons in each chart of the figure. No significant difference in tissue metal accumulation could be discerned between top position oysters at GNPMS and top positioned oysters at KBCCM, or between bottom position oysters at GNPMS and bottom positioned oysters at KBCCM (data not shown). The various tissues of the one-year old Jamaica Bay oysters readily accumulated arsenic, cadmium, copper and zinc. The metal content of the tissues were not homogeneously distributed and were different for the oysters maintained at the top as compared to the bottom. Soft tissues accumulated metals in the µg/g dwt range,

which is comparable to other published reports for *C. virginica* growing in other areas^{34,35}. Of the four metals tested, zinc showed the most bioaccumulation in all oyster tissues including shell.

In soft tissues, zinc accumulations were greater than copper, which were greater than cadmium. Arsenic showed the lowest soft tissue accumulations. For arsenic, cadmium and copper, the gills, mantle and palps accumulated the most metals; while for copper, the stomach accumulated considerable more than the other tissues. With all four metals, the soft tissues of oysters positioned one foot below the surface accumulated more metal than their bottom positioned counterpart, with the exception of stomach which not only showed very high copper accumulations but demonstrated a greater copper accumulation in bottom dwelling oysters.

Shell tissue also accumulated metals but to a lesser extent. With the exception of zinc, shell metal accumulations were 10-100x less than metal accumulations seen in any one type of oyster tissue. Zinc and copper accumulation were in the $\mu\text{g/g}$ dwt range while cadmium and arsenic accumulations were in the ng/g dwt range. Average shell zinc levels ($\mu\text{g/g}$) were higher in top positioned oysters (top: 123 ± 20 , bottom: 47 ± 16); average shell cadmium levels (ng/g) were higher in top positioned oysters (top: 28 ± 9 , bottom: 12 ± 4); and average shell arsenic levels (ng/g) were higher in top positioned oysters (top: 213 ± 37 , bottom: 92 ± 18). Only copper showed slightly higher average accumulations ($\mu\text{g/g}$) in bottom positioned oysters (top: 1.033 ± 0.145 , bottom: 1.450 ± 0.352).

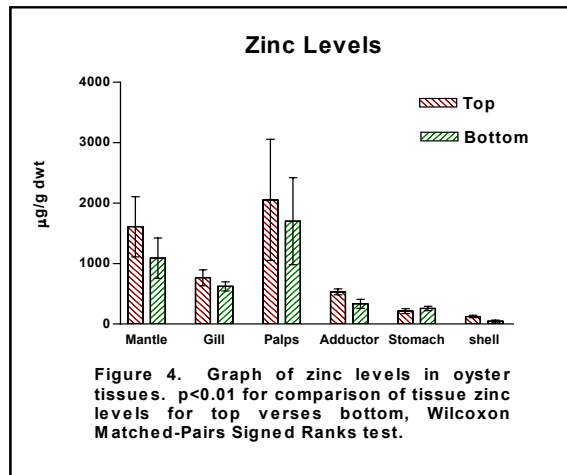
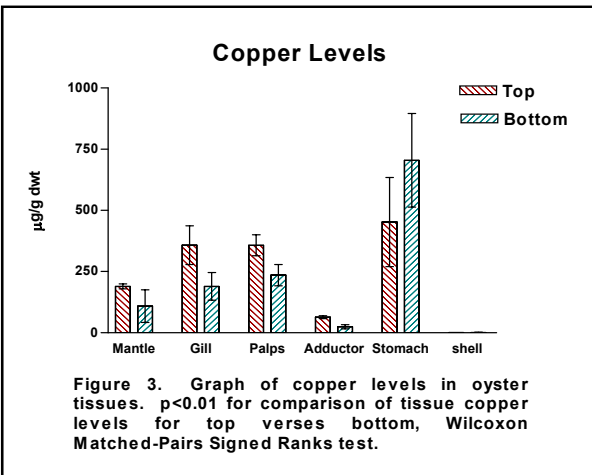
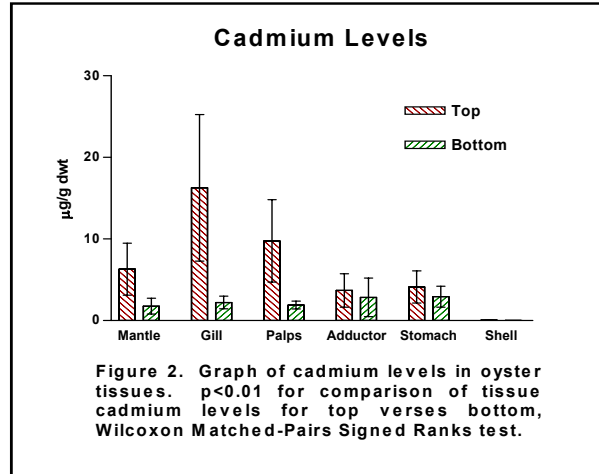
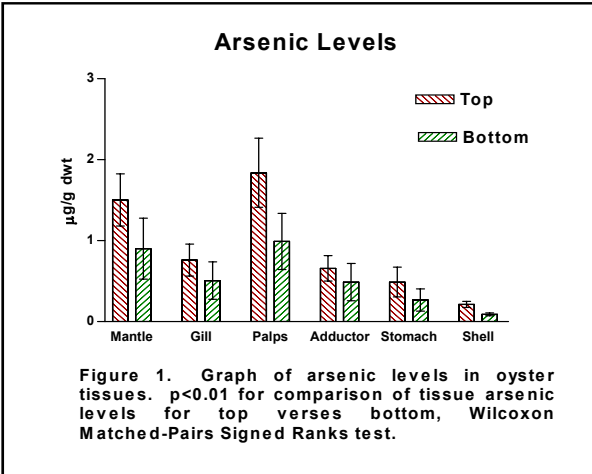
Discussion

Marine bivalves are filter feeders that take up and accumulate metals and other pollutants from the water column or via ingestion of contaminants adsorbed to phytoplankton, detritus and sediment particles. Because they are sessile, they reflect local contaminant concentrations more

accurately than crustaceans and free-swimming finfish. Marine bivalves such as oysters and mussels have been extensively used as model organisms in environmental studies of water quality^{36,37}. Trace metals are taken up and accumulated by oysters and many other marine invertebrates to tissue and body concentrations usually much higher on a wet weight basis than concentrations in the surrounding seawater³⁸.

In this study we measured the one year bioaccumulation and tissue distribution of four metal pollutants (arsenic, cadmium, copper and zinc) in *C. virginica* seed, that had been transplanted to the bay in 2002 at two different depths: one foot below the surface and one foot above the sediment. Past years of unregulated industrial, rural and residential dumping, and previously inadequate wastewater treatment has caused Jamaica Bay sediment to be a major reservoir of metal pollutants. While the content and degree of metal contamination in the sediment can vary greatly depending upon site tested, the presence of many metal pollutants throughout Jamaica Bay sediment is widespread. In a 2002 report to the New York State Department of Environmental Conservation and the NY/NJ Port Authority¹³, sediment sampled at 3 different sites in Jamaica Bay indicated that the content of many metals including arsenic, cadmium, copper and zinc were present at levels above the "Effects Range Low Level" (ERL) guideline and at some sites were above the "Probable Effects Level" (PEL) guideline. Arsenic levels ranged from 8.6-11 mg/kg dwt. (ERL and PEL, 8.2 and 41.6, respectively). Cadmium levels ranged from 2.2-5.7 mg/kg dwt. (ERL and PEL, 1.2 and 4.2, respectively). Copper levels ranged from 85-160 mg/kg dwt. (ERL and PEL, 34 and 108.2, respectively). Zinc levels ranged from 45-307 mg/kg dwt. (ERL and PEL, 150 and 271, respectively). A 1991 report on the annual input of heavy metals to Jamaica Bay revealed that sewage effluent carries the largest quantities of zinc, copper and cadmium to the bay³⁹.

Our results indicate that over a one-year period of growth in Jamaica Bay,



considerable amounts of arsenic, cadmium, copper and zinc can accumulate in the tissues of oyster seed. Arsenic and cadmium are believed to serve no essential role in animal or plant metabolism^{40,41}. Copper and zinc function as micronutrients, but when in excess even these essential elements can cause ecotoxicological effects. Therefore, organisms that are exposed to and accumulate metal pollutants require a means of physiological detoxification, typically by binding to high affinity sites in inorganic granules, or to various proteins like apoferritin or cysteine-rich metallothionein^{42,43}.

Arsenic is the twentieth most abundant element in the earth's crust and sources of aquatic arsenic are both natural as well as anthropogenic. It is released naturally to the atmosphere from volcanic eruptions

and forest fires⁴⁴ and to water via natural weathering processes⁴⁵. Anthropogenic sources include pollution due to fossil fuel combustion, mining/smelting, effluents from sewage treatment facilities, leaching from hazardous waste disposal sites, and production and use of arsenic compounds as pesticides and as a wood preservative⁴⁶. The toxicity of arsenicals is highly dependent upon the nature of the compound with trivalent arsenic compounds considered more toxic than pentavalent arsenic compounds, and inorganic forms more toxic than organic forms^{47,48}. Arsenic and arsenic-containing organic compounds have not been shown to accumulate to any great extent in aquatic organisms⁴⁹, with perhaps the exception of *C. virginica*, which in a bioaccumulation test was able to generate a very high bioconcentration

factor (BCF) of 350 after 112 days of exposure to phytoplankton containing arsenic⁵⁰. In addition *C. virginica* may be particularly resistant to the toxic effects of arsenic for data comparing the acute toxicity of inorganic arsenic (III) indicated a very high acute value for *C. virginica* compared to *C. gigas*, a related species (7500 ug/l; 326 ug/l, respectively)⁵¹.

Cadmium, a trace metal widely distributed in soil, air, water, and living things⁴⁰, is naturally present in zinc, lead, and copper deposits. It is also released into the environment from several anthropogenic sources including smelting and mining, electroplating, application of phosphate fertilizers, waste disposal operations, and the manufacturing and disposal of paints, alloys, batteries, and plastics^{52,53,54,55,56}. Cadmium is a cumulative human toxicant that can cause a variety of adverse health problems including renal dysfunction and degenerative bone disease^{57,58}. Cadmium has been found to bioaccumulate in fish and shellfish tissues in estuarine/marine waters^{59,60} nationwide and New York has issued advisories for this metal in all of its marine coastal waters.

Copper is plentiful in the environment and is an essential micronutrient for the normal growth and metabolism of all living organisms^{61,62,63}. The United States is the major world producer and consumer of copper and its compounds. Copper releases to the global biosphere—which may approach 1.8 million metric tons per year⁶⁴ - come mostly from anthropogenic activities⁶⁵. Copper is among the most toxic of the heavy metals in freshwater and marine biota^{61,66}. Excess copper can cause cellular damage by generating oxygen free radicals and inactivating biological thiols into disulfides⁶². Inputs of copper into aquatic ecosystems have increased sharply during the past century due to a number of reasons including atmospheric fallout from industrial activities, waste and industrial discharges, and leaching of antifouling marine paints and wood preservatives^{67,68}. Copper is elevated in sediments of many marinas, probably as a result of copper containing antifouling paints used on boats housed in these marinas⁶⁹.

Bioavailability and toxicity of copper to aquatic organisms depends on the total concentration of copper and its speciation⁷⁰. In marine ecosystems, the high copper levels measured in heavily contaminated coastal areas sometimes approach the incipient lethal concentrations for some organisms⁷¹. Among marine organisms, the highest accumulations are generally found in molluscan tissues and soft parts, especially those of cephalopods and oysters⁶⁵. Bioconcentration factors for copper are highest for American oysters after exposure for 140 days (20,700-28,200), and lowest for bay scallops (*Argopecten irradians*) after exposure for 112 days (3,310) and for softshell clams after exposure for 35 days (3,300)⁷².

Zinc is ubiquitous in the tissues of plants and animals⁷³ and functions as a cofactor for many enzymes in both aquatic and terrestrial organisms⁷⁴. Like many other trace metals, zinc can be toxic if present at high concentrations in aquatic ecosystems^{75,76,77}. Most of the zinc introduced into aquatic environments eventually is partitioned into the sediments. Anthropogenic sources include domestic and industrial sewage, combustion of fossil fuels and solid wastes, corrosion of zinc alloys and galvanized surfaces, road surface runoff, smelting and mining operations, and erosion of agricultural soils^{78,79,80,81,82}. In seawater zinc exists in a dissolved state, as a solid precipitate, or adsorbed to particle surfaces. Zinc in molluscs is usually associated with high molecular weight proteins, with diet (as opposed to ambient water zinc concentrations), from collection locales with elevated sediment zinc burdens, and with particulate matter from dredging and storm perturbations^{83,84}. Molluscan life processes do not seem to be affected by excess zinc accumulations and zinc is frequently accumulated far in excess of the organism's immediate needs⁸³.

In this study *C. virginica* seed, transplant to Jamaica Bay, readily accumulated arsenic, cadmium, copper and zinc and the distribution of these metals were not homogeneous throughout oyster

tissues. The unequal distribution of all 4 metals among the tested oyster tissue is not likely a random event. Copper values were particularly high in stomach tissue and are more likely a reflection of the copper-laden microalgae and detritus the animal ingested rather than actual tissue accumulations. For the other tissues, the cellular concentration of metallothionein and other metal binding organic or inorganic compounds probably varies depending upon physiological parameters and each tissue's role in metal detoxification. In the case of copper and zinc, which are micronutrients, the unequal distribution and known extensive accumulations possible in the American oyster may correlate with some physiological function for these two metals. Both of these nutrients are exclusively sequestered in oyster amebocytes, a cell type credited with many indispensable responsibilities for oyster survival, including antimicrobial activities for defense and nutrition⁸⁵. The fact that copper and zinc accumulations, which were most abundant in the gills, mantle and palps, were not homogeneous throughout oyster tissue may be more associated with amebocyte function and distribution than a detoxified storage of these metals.

A number of studies have reported on metal accumulations in the shell of *C. virginica* from various sites^{86,87,88,89}. Our results are in accordance with previous findings with shell tissue accumulating metals but to a lesser extent. Copper and zinc concentrations were in the ug/g dwt range, while cadmium and arsenic concentrations were in the ng/g range. Even though metal accumulations in shell were much lower than that found in oyster soft tissue, if one considers that the shell of the animal typically makes up over 75% of the animal's weight, then the overall amount of metal present in the shell of the whole animal could be significant⁸⁷. Therefore measurements of metal accumulations in oyster shell are important for they can represent a terrestrial reservoir of aquatic pollutants that can be adsorbed or released into the environment depending upon the nature of the metal and

ambient concentrations. In addition, metals and other aquatic pollutants may influence the normal mineralization and metal composition of oyster shells, possibly decreasing shell strength and stability. Frazier⁹⁰ believed that shell thinning was linked to effects of copper and cadmium on shell calcification enzymes. If competent shell structure depends on metal composition, then deviations in metal composition may adversely affect shell quality, compromising the long-term health and survival of the animal.

Comparing metal accumulations in *C. virginica* seed grown at different depths in the bay, our results indicate that for most of the tissues studied, top-maintained oysters accumulated more metal pollutants than oysters maintained near the sediment. Differences in general water quality parameters such as pH, temperature, dissolved O₂, salinity, and turbidity could not account for our findings because these parameters, which were monitored throughout the experimental period, indicated no statistical variation at either site or at either depth¹⁹. Our results also indicate that wide variations in metal accumulations were possible even among individuals positioned at the same site and same depth. This is commonly reported by monitoring agencies for bioindicator organisms and can be due to many causes including age, seasonal factors and individual variability within the population³⁴. While our oysters were all of the same age and reproductive status (subadult), differences in metal accumulations could still exist due to feeding patterns, extent of glycogen stores, adaptability to acute stresses, or overall health of the organism^{91,92,93}. Even sex can be a factor for variations in metal accumulations for in the bivalve *Donax trunculus* L., females were shown to accumulate higher concentrations of zinc than males⁹⁴. Differences in feeding patterns or overall health may also be an explanation for the disparity in metal accumulations between surface and bottom dwelling oysters. Bivalves accumulate pollutants not only from the water column and

metal-laden detritus, but also via ingestion of metal contaminated phytoplankton. While our measurements of chlorophyll-a in surface/bottom water showed no consistent variation¹⁹, the extent of feeding and the species and metal content of the microalgae available to the oysters near the surface might have been different from that available to oysters growing near the sediment. It also is possible that the increase in metal accumulations seen in top maintained oysters may have been due to a difference in top/bottom overall health or early infection rates with known oyster pathogens such as *Haplosporidium nelsoni* (MSX disease) or *Perkinsus marinus* (Dermo disease). While neither organism is harmful to humans, nor a health threat to humans who ingest infected shellfish, these parasites can chronically weaken and eventually kill *C. virginica* over a period of years^{95,96}. Other studies have shown a positive correlation between tissue burden of certain metals, especially copper and zinc, and immuno-defense related characteristic of oysters^{85,97,98}. Also to consider is the fact that top but not bottom dwelling oysters may have been exposed to additional bay contaminants, including floatables and organic pollutants, that either enhanced metal accumulations or compromised oyster health and therefore soft tissue growth. Differences in soft tissue weights can significantly affect trace metal concentrations by simply diluting or concentrating the animal's total metal body burden^{99,100,101,102,103,104,105}. It was a surprise that oysters positioned near the sediment not only had lower overall metal accumulations, but also had better growth rates when compared to top grown oysters¹⁹. It is currently unknown whether this pattern of greater growth and lesser metal accumulations in Jamaica Bay is site specific or if metal accumulation would have been greater and growth worse in bottom dwelling oysters if a more polluted site were available for our study. The fact that the bottom placed oysters accumulated less tissue arsenic, cadmium, copper and zinc than their top placed counterpart is interesting and may itself be an explanation for the faster bottom

growth rate. However, only 5 tissues and these 4 metal pollutants were studied in this report. More work needs to be done here and at additional sites in Jamaica Bay to determine if other metal pollutants and other oyster tissues show similar patterns of metal accumulations.

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References

- ¹United States Fish and Wildlife Service (USFWS). 1997. Significant Habitats and Habitat Complexes of the New York Bight Watershed, Southern New England–New York Bight Coastal Ecosystems Program, Jamaica Bay and Breezy Point, Complex #16, November 1997.
- ²MacKenzie, C.L., Jr., V.G. Burrell, Jr., A. Rosenfield and W.L. Hobart (editors). 1997. The history, present condition, and future of the molluscan fisheries of North and Central America and Europe, vol. 1. Atlantic and Gulf Coasts. US Dept. Commerce, NOAA Tech. Rep. NMFS 127, Seattle, WA, 234 pp.
- ³New York/New Jersey Harbor Estuary Program (NYNJHEP). 2004. *Health of the Harbor - The first Comprehensive Look at the State of the NY/NJ Harbor Estuary*. Hudson River Foundation, NY pp 82.
- ⁴Franz, D.R., 1982. An historical perspective on mollusks in lower New York Harbor, with emphasis on oysters, p. 181-197. In G. F. Mayer (ed.), *Ecological Stress and the New York Bight: Science and Management*. Estuarine Research Federation, Columbia, South Carolina. 715 p.
- ⁵Rothschild, B.J., J.S. Ault, P. Gouletquer, and M. Heral, 1994. Decline of the Chesapeake Bay oyster population: a century of habitat destruction and overfishing. *Mar. Ecol. Prog. Ser.* **111**: 29-39.
- ⁶Woods, H., W.J. Hargis, Jr., C.H. Hershner and P. Mason, 2005. Disappearance of the natural emergent 3-dimensional oyster reef system of the James River, Virginia, 1871-1948. *J. Shellfish Res.* **24**:139-142.
- ⁷MacKenzie, C.L., 1996a. History of oystering in the United States and Canada, featuring the eight greatest oyster estuaries. *Mar. Fish. Rev.* **58**: 1-78.

- ⁸United States Army Corps of Engineers (USACE), 2000. Reconnaissance Study, Hudson-Raritan Estuary Environmental Restoration Study, Section 905(b) WRDA 86 Preliminary Analysis, New York, June 2000. pp 38.
- ⁹Ramondetta, P.J and W.H. Harris, 1978. Heavy metals distribution in Jamaica Bay sediments. *Environmental Geology* Volume **2(3)**: 145-149.
- ¹⁰Franz, D.R. and W.H. Harris, 1985. Benthos Study. Jamaica Bay Wildlife Refuge, Gateway National Recreation Area, Brooklyn, New York. Final Report, DOI, Contract #CX 1600-I-0031.
- ¹¹Staubitz, W.W. and S.W. Wolcott, 1985. Hydraulic and sediment characteristics at the North Channel Bridge, Jamaica Bay, New York, U.S. Geological Survey Water-Resources Investigations Report 85-4085, 43 p.
- ¹²Adams, D.A., J.S. O'Connor and S.B. Weisberg, 1998. Final Report: Sediment Quality of the NY/NJ Harbor System- An Investigation under the Regional Environmental Monitoring and Assessment Program (REMAP). EPA/902-R-98-001. USEPA-Region 2, Division of Environmental Science and Assessment. Edison, NJ. www.epa.gov/emap/remap/html/docs/nynjharbor.html
- ¹³New York State Department of Environmental Conservation (NYSDEC) and the NY/NJ Port Authority (NYNJPA), 2002 Report on Bioassay analyses conducted on sediments collected from Jamaica Bay, Far Rockaway, New York June 2002. Prepared by Barry A. Vittor & Associates, Inc. Mobile, AL, September 2002.
- ¹⁴Adams, D.A. and S. Benyi, 2003. Final Report: Sediment Quality of the NY/NJ Harbor System: A 5 year revisit 1993/4 –1998, An Investigation under the Regional Environmental Monitoring and Assessment Program (REMAP). EPA/902-R-03-002. USEPA-Region 2, Division of Environmental Science and Assessment. Edison, NJ. http://www.epa.gov/emap/remap/html/docs/NY_NJ_Harbor98.pdf.
- ¹⁵Ortiz, M.T., A.M. Stavroulakis, A. Love, P. Lanzetta and P. Pilchman, 2005. Characterization of Potential Transplantation Sites for Eelgrass (*Zostera marina* L.) in Jamaica Bay, New York and Eelgrass Growth in a Laboratory Microcosm Mimicking Field Conditions. *In Vivo* **26(3)**: 4-18.
- ¹⁶Wikfors, G. H., J. Twarog, W. Joseph, G. E. Ferris, B. C. Smith and R. Ukeles, 1994. Survival and growth of post-set oysters and clams on diets of cadmium-contaminated microalgal cultures. *Mar. Environ. Res.* **37(3)**: 257-281.
- ¹⁷Keppler, C. and A. H. Ringwood, 2001. Expression of P-glycoprotein in the gills of oysters, *Crassostrea virginica*: seasonal and pollutant related effects. *Aquat. Toxicol.* **54(3-4)**:195-204.
- ¹⁸Gifford, S.P., G.R MacFarlane, W.A. O'Connor and R.H. Dunstan, 2006. Effect of the pollutants lead, zinc, hexadecane and octocosane on total growth and shell growth in the akoya pearl oyster, *Pinctada imbricate*. *Journal of Shellfish Research* **25(1)**: 159-165.
- ¹⁹Sarinsky, G., M.A. Carroll, E. Nduka and E.J. Catapane, 2005. Growth and Survival of the American Oyster *Crassostrea virginica* in Jamaica Bay, New York. *In Vivo* **27(1)**: 15-26.
- ²⁰Cunningham, P.A., 1979. The use of bivalve molluscs in heavy metal pollution research: *In Marine Pollution: Functional Responses*. W.B. Vernberg, F.P. Thurberg, A. Calabrese and F.J. Vernberg (eds.) Academic Press, NY.
- ²¹Bourgoin B.P., 1990. *Mytilus edulis* shell as a bioindicator of lead pollution: Considerations on bioavailability and variability. *Mar. Ecol. Prog. Ser.* **61**: 253-262.
- ²²Phillips, D. and P. Rainbow, 1993. *Biomonitoring of trace aquatic contaminants*. London, New York: Elsevier Applied Science. pp 371.
- ²³Boening, D.W., 1999. An evaluation of bivalves as biomonitors of heavy metal pollution in marine waters. *Environ. Monit. Assess.* **55**: 459-470.
- ²⁴U.S. EPA, 1993a. Vol 1 Fish Sampling and Analysis, 3rd edition, Ch.3, p16.
- ²⁵U.S. EPA 1993b. Vol 1 Fish Sampling and Analysis, 3rd edition, Ch.4, p3.
- ²⁶NOAA, 1998 (on-line). "Chemical Contaminants in Oysters and Mussels" by Tom O'Connor.
- ²⁷Capar, S.G. and N.J. Yess, 1996. US Food and Drug Administration survey of cadmium, lead and other elements in clams and oysters. *Food Addit. Contam* **13(5)**: 553-560.
- ²⁸Bu-Olayan, A.H. and M.N. Subrahmanyam, 1997. Accumulation of copper, nickel, lead and zinc by snail, *Lunella coronatus* and pearl oyster, *Pinctada radiata* from the Kuwait coast before and after the Gulf War oil spill. *Sci Total Environ* **197(1)**: 161-165.
- ²⁹Spooner, D.R., W. Maher and N. Otway, 2003. Trace Metal Concentrations in Sediments and Oysters of Botany Bay, NSW, Australia. *Archives of Environmental Contamination and Toxicology* **45(1)**: 0092-0101.
- ³⁰Scanes, P.R. and A.C. Roach, 1999. Determining natural background concentrations of trace metals in oysters from New South Wales, Australia *Environ Poll* **105(3)**: 437-446.
- ³¹Abbe, G.R., G.F. Riedel and J.G. Sanders, 2000. Factors that influence the accumulation of copper and cadmium by transplanted eastern oysters (*Crassostrea virginica*) in the Patuxent River, Maryland. *Marine Environmental Research* **49(4)**: 377-396.
- ³²Fang, Z.Q., R.Y.H. Cheung and M.H. Wong, 2001. Heavy Metal concentrations in edible bivalves and gastropods available in the major markets of the Pearl River Delta. *J Environ Sci* **13(2)**: 210-217.
- ³³Luckenbach, M.W., F.X. O'Beirn and J. Taylor, 1999. *An Introduction to Culturing Oysters in Virginia*, Virginia Institute of Marine Science Press, Gloucester Point, VA. pp24. On-line at <http://www.vims.edu/oystergarden/CulturingOysters.pdf>.

- ³⁴Daskalakis, K.D., 1996. Variability of metal concentrations in oyster tissue and implications to biomonitoring Marine Pollution Bulletin **32(11)**: 794-801.
- ³⁵Karouna-Renier, N.K., R.A. Snyder, J.G. Allison, M.G. Wagner and K.R. Rao, 2007. Accumulation of organic and inorganic contaminants in shellfish collected in estuarine waters near Pensacola, Florida: Contamination profiles and risks to human consumers. Environmental Pollution **145(2)**: 474-488.
- ³⁶Wang, W.X., N.S. Fisher and S.N. Luoma, 1996. Kinetic determinations of trace element bioaccumulation in the mussel *Mytilus edulis*. Mar Ecol Prog Ser **140**: 91-113.
- ³⁷Griscom, S.B., N.S. Fisher and S.N. Luoma, 2000. Geochemical Influences on Assimilation of Sediment-Bound Metals in Clams and Mussels. Environ. Sci Technol **34**: 91-99.
- ³⁸Rainbow, P.S., 1993. The Significance of Trace Metal Concentrations in Marine Invertebrates, In: *Ecotoxicology of Metals in Invertebrates*. Proceedings of a session at the First Society of Environmental Toxicology and Chemistry-Europe Conference held in Sheffield, England, 7-10 April 1991. Lewis Publishers, Boca Raton, Florida, 1993. p 3-23. 8 tab, 58 ref.
- ³⁹Seidemann, D., 1991. Metal pollution in sediments of Jamaica Bay, New York, USA - An urban estuary, Environmental Management **15(1)**: 73-81.
- ⁴⁰Pinot, F., S.E. Kreps, M. Bachelet, P. Hainaut, M. Bakonyi and B.S. Polla, 2000. Cadmium in the environment: sources, mechanisms of biotoxicity, and biomarkers. Rev. Environ. Health **15**: 299-323.
- ⁴¹Satarug, S., M. Nishijo, J.M. Lasker, R.J. Edwards and M.R. Moore, 2006. Kidney dysfunction and hypertension: role for cadmium, P450 and heme oxygenases? Tohoku J. Exp. Med. **208**: 179-202.
- ⁴²Rainbow, P.S., 1997. Ecophysiology of Trace Metal Uptake in Crustaceans. Estuarine, Coastal and Shelf Science **44**: 169-175.
- ⁴³Roesijadi, G., 1996. Metallothionein and its role in toxic metal regulation. Comp. Biochem. Physiol. C **113**: 117-123.
- ⁴⁴Walsh, P.R., R.A. Duce and J.L. Fasching, 1979. Considerations of the enrichment, sources, and flux of arsenic in the troposphere. J. Geophys Res. **84(C4)**: 1719-1726.
- ⁴⁵U.S. EPA (U.S. Environmental Protection Agency). 1982. Arsenic. In: *Intermedia Priority Pollutant Guidance Documents*. Office of Pesticides and Toxic Substances, Washington, DC.
- ⁴⁶Eisler, R., 1988. Arsenic Hazards to Fish, Wildlife, and Invertebrates: A Synoptic Review. Biological Report 85(1.12), Contaminant Hazard Reviews, Report No. 12. Fish and Wildlife Service, U.S. Department of the Interior, Laurel, MD.
- ⁴⁷Vallee, B.L., D.D. Ulmer and W.E.C. Wacker, 1960. Arsenic toxicology and biochemistry. Arch. Ind. Health **21**: 132-151.
- ⁴⁸Penrose, W.R., 1974. Arsenic in the marine and aquatic environments: Analysis, occurrence and significance. C.R.C. Crit. Rev. in Environ. Contam. **4**: 465-482.
- ⁴⁹National Academy of Sciences (NAS), 1977a. *Arsenic*. Committee on Medical and Biologic Effects of Environmental Pollutants, National Research Council, Washington, DC.
- ⁵⁰Zarogian, G.E. and G.L. Hoffman, 1982. Arsenic uptake and loss in the American oyster, *Crassostrea virginica*. Environ. Monit. Assessment **1**: 345-358.
- ⁵¹EPA 1985. Ambient Water Quality Criteria for Arsenic - 1984. EPA 440/5-84-033, January 1985.
- ⁵²May, T.W. and G.L. McKinney, 1981. Cadmium, lead, mercury, arsenic and selenium concentrations in freshwater fish, 1976-1977 - National Pesticide Monitoring Program. Pesticides Monitoring Journal **15(1)**: 14-38.
- ⁵³Farag, A.M., D.F. Woodward, J.N. Goldstein, W. Brumbaugh and J.S. Meyer, 1998. Concentrations of metals associated with mining waste in sediments, biofilm, benthic macroinvertebrates, and fish from the Coeur D'Alene River Basin, Idaho. Arch. Environ. Contam. Toxicol. **34**: 119-127.
- ⁵⁴U.S. EPA (U.S. Environmental Protection Agency), 1978. Metal Bioaccumulation in Fish and Aquatic Invertebrates. EPA-600/3-78-103. Environmental Research Laboratory, Office of Research and Development, Springfield, VA.
- ⁵⁵U.S. EPA (U.S. Environmental Protection Agency), 1979. Health Assessment Document for Cadmium. EPA-600/8-79-003. Environmental Standards and Criteria, Office of Research and Development, Research Triangle Park, NC.
- ⁵⁶U.S. EPA (U.S. Environmental Protection Agency), 1987. Cadmium Health Advisory Draft. Office of Drinking Water, Washington, DC.
- ⁵⁷Goering P.L., M.P. Waalkes and C.D. Klaassen, 1995. Toxicology of cadmium, in *Toxicology of Metals* (Goyer RA and Cherian MG eds) pp 190-214, Springer-Verlag, Berlin, Germany.
- ⁵⁸Waisberg, M., P. Joseph, B. Hale and D. Beyersmann, 2003. Molecular and cellular mechanisms of cadmium carcinogenesis. Toxicology **192**: 95-117.
- ⁵⁹NOAA (National Oceanic and Atmospheric Administration), 1987. National Status and Trends Program for Marine Environmental Quality-Progress Report: A Summary of Selected Data on Chemical Contaminants in Tissues Collected During 1984, 1985 and 1986. NOAA Technical Memorandum NOS OMA 38. U.S. Department of Commerce, Rockville, MD.
- ⁶⁰NOAA (National Oceanic and Atmospheric Administration), 1989. National Status and Trends Program for Marine Environmental Quality - Progress Report: A Summary of Selected Data on Tissue Contamination from the First Three Years (1986-1988) of the Mussel Watch Project. NOAA Technical Memorandum NOS OMA 49. U.S. Department of Commerce, Rockville, MD.
- ⁶¹Schroeder, H.A., A.P. Nason, I.H. Tipton and J.J. Balassa, 1966. Essential trace metals in man: copper. Journal of Chronic Diseases **19**: 1007-1034.

- ⁶²Aaseth, J. and T. Norseth, 1986. Copper. Pages 233-254 in L. Friberg, G.F. Nordberg, and V.B. Vouk, editors. *Handbook on the toxicology of metals*. Second edition. Volume II: specific metals. Elsevier, New York.
- ⁶³Carbonell, G. and J.V. Tarazona, 1994. Toxicokinetics of copper in rainbow trout (*Oncorhynchus mykiss*). *Aquatic Toxicology* **29**: 213-221.
- ⁶⁴National Academy of Sciences (NAS), 1977b. Copper. Committee on Medical and Biologic Effects of Environmental Pollutants, National Research Council, National Academy of Sciences, Washington, D.C. 115 pp.
- ⁶⁵Eisler, R., 1998. Copper hazards to fish, wildlife, and invertebrates: a synoptic review. U.S. Geological Survey, Biological Resources Division, Biological Science Report USGS/BRD/BSR--1998-0002.
- ⁶⁶Betzer, S.B. and P.P. Yevich, 1975. Copper toxicity in *Busycon canaliculatum* L. *Biological Bulletin* **148**:16-25.
- ⁶⁷Nriagu, J.O., 1979a. The global copper cycle. Pages 1-17 in J. O. Nriagu, editor. *Copper in the environment*. Part 1: ecological cycling. John Wiley, N.Y.
- ⁶⁸Nriagu, J.O., 1979b. Copper in the atmosphere and precipitation. Pages 45-75 in J.O. Nriagu, editor. *Copper in the environment*. Part 1: ecological cycling. John Wiley, NY.
- ⁶⁹Hall, W.S., S.J. Bushong, L.W. Hall, Jr., M.S. Lenkevich and A.E. Pinkney, 1988. Monitoring dissolved copper concentrations in Chesapeake Bay, U.S.A. *Environmental Monitoring and Assessment* **11**: 33-42.
- ⁷⁰Hung, T.C., A. Chuang, S.J. Wu and C.C. H. Tsai, 1990. Relationships among the species and forms of copper and biomass along the Erhjin Chi coastal water. *Acta Oceanographic Taiwanica* **25**: 65-76.
- ⁷¹Neff, J.M. and J.W. Anderson, 1977. The effects of copper (II) on molting and growth of juvenile lesser blue crabs *Callinectes similis*. Pages 155-165 in C. S. Giam, editor. *Pollutant effects on marine organisms*. D.C. Heath and Company, Lexington, Massachusetts.
- ⁷²U.S. Environmental Protection Agency (USEPA), 1980. Ambient water quality criteria for copper. U.S. Environmental Protection Agency Report 440/5-80-036. 162 pp.
- ⁷³Rosser, B.W.C. and J.C. George, 1986. Molt-induced muscle atrophy decreases the zinc content of the pectoralis of the giant Canada goose (*Branta canadensis maxima*). *Experientia* **42**: 549-550.
- ⁷⁴Coombs, T.L., 1972. The distribution of zinc in the oyster *Ostrea edulis* and its relation to enzymic activity and to other metals. *Mar. Biol.* **12**: 170-178.
- ⁷⁵Eisler, R. and G.R. Gardner, 1973. Acute toxicology to an estuarine teleost of mixtures of cadmium, copper and zinc salts. *J. Fish Biol.* **5**: 131.
- ⁷⁶Eisler, R., 1993. Zinc hazards to fish, wildlife, and invertebrates: a synoptic review. U.S. Fish and Wildlife Service Biological Report 10. 106 pp.
- ⁷⁷Eisler, R. and M. Wapner, 1975. Second annotated bibliography on biological effects of metals in aquatic environments. U.S. Environmental Protection Agency, Office of Research and Development, Environmental Research Laboratory, Narragansett, RI. EPA-600/3-75-008.
- ⁷⁸Weatherley, A.H., P.S. Lake and S.C. Rogers, 1980. Zinc pollution and the ecology of the freshwater environment. Pages 337-417 in J. O. Nriagu, editor. *Zinc in the environment*. Part I: ecological cycling. John Wiley, NY.
- ⁷⁹Spear, P.A., 1981. Zinc in the aquatic environment: chemistry, distribution, and toxicology. National Research Council of Canada Publication NRCC 17589. 145 pp.
- ⁸⁰Mirenda, R.J., 1986. Acute toxicity and accumulation of zinc in the crayfish, *Orconectes virilis* (Hagen). *Bulletin of Environmental Contamination and Toxicology* **37**: 387-394.
- ⁸¹Llobet, J.M., J.L. Domingo, M.T. Colomina, E. Mayayo and J. Corbella, 1988a. Subchronic oral toxicity of zinc in rats. *Bulletin of Environmental Contamination and Toxicology* **41**: 36-43.
- ⁸²Buhl, K.J. and S.J. Hamilton, 1990. Comparative toxicity of inorganic contaminants related by placer mining to early life stages of salmonids. *Ecotoxicology and Environmental Safety* **20**: 325-342.
- ⁸³Eisler, R., 1981. *Trace metal concentrations in marine organisms*. Pergamon Press, NY, 687 pp.
- ⁸⁴Eisler, R., 1984. Trace metal changes associated with age of marine vertebrates. *Biological Trace Element Research* **6**: 165-180.
- ⁸⁵Fisher, W.S., 2004. Antimicrobial activity of copper and zinc accumulated in Eastern oyster amebocytes. *Journal of Shellfish Research* **23(2)**: 321-351.
- ⁸⁶Smith, R.A. and E.R. Wright, 1962. Elemental composition of oyster shell. *Texas J. Sci.* **14**: 222-224.
- ⁸⁷Wolfe, D.A., 1970. Levels of stable Zn and ⁶⁵Zn in *Crassostrea virginica* from North Carolina. *J. Fish. Res. Bd. Canada* **27**: 47-57.
- ⁸⁸Windom, H.L. and R.G. Smith, 1972. Distribution of iron, manganese, copper, zinc, and silver in oysters along the Georgia coast. *J. Fish. Res. Bd. Canada* **29**: 450-452.
- ⁸⁹Ferrell, R.E., T.E. Carville and J.D. Martinez, 1973. Trace metals in oyster shells. *Environ. Lett.* **4**: 311-316.
- ⁹⁰Frazier, J.M., 1976. The dynamics of metals in the American oyster, *Crassostrea virginica*. II. Environmental effects. *Chesapeake Sci.* **17**: 188-197.
- ⁹¹Phillips, D.J.H., 1976. The common mussel *Mytilus edulis* as an indicator of pollution by zinc, cadmium, lead and copper. 1. Effects of environmental variables on uptake of metals. *Mar. Biol.* **38**: 71-80.
- ⁹²Popham, J.D., J.M. D'Auria, 1983. Combined effect of body size, season, and location on trace element levels in mussels (*Mytilus edulis*). *Archs environ. Contam. Toxicol* **12**: 1-14.

- ⁹³Martin, N., G. Ichikawa, J. Goetzi, M. de Reyes and M.D. Stephenson, 1984. Relationships between physiological stress and trace toxic substances in the bay mussel, *Mytilus edulis* from San Francisco Bay, California. *Mar. Environ. Res.* **11**: 91-110.
- ⁹⁴Marina, M. and O. Enzo, 1983. Variability of zinc and manganese concentrations in relation to sex and season in the bivalve *Donax trunculus* Marine Pollution Bulletin **14(9)**: 342-346.
- ⁹⁵Ford, S.E. and M.R. Tripp, 1996. Diseases and Defense Mechanisms. In: Kennedy, V.S., R.I.E. Newell and A.F. Eble (eds), *The Eastern Oyster Crassostrea virginica*. Maryland Sea Grant College, College Park, Maryland, pp. 581-660.
- ⁹⁶Lenihan, H.S., F. Micheli, S.H. Shelton and C.H. Peterson, 1999. The influence of multiple environmental stressors on susceptibility to parasites: an experimental determination with oysters. *Limnology and Oceanography*. **44(3)**: 910-924.
- ⁹⁷Fisher, W.S., L.M. Oliver, J.T. Winstead and E.R. Long, 2000. A survey of oysters *Crassostrea virginica* from Tampa Bay, Florida: associations of internal defense measurements with contaminant burdens. *Aquatic Toxicology* **51(1)**: 115-138.
- ⁹⁸Oliver, L.M., W.S. Fisher, J.T. Winstead, B.L. Hemmer and E.R. Long, 2001. Relationships between tissue contaminants and defense-related characteristics of oysters (*Crassostrea virginica*) from five Florida bays. *Aquatic Toxicology* **55(3,4)**: 203-222.
- ⁹⁹Simpson, R.D., 1979. Uptake and loss of zinc and lead by mussels (*Mytilus edulis*) and relationships with body weight and reproductive cycle. *Mar. Pollut. Bull.* **10**: 74-78.
- ¹⁰⁰Strong, C.R., S.N. Luoma, 1981. Variations in the correlation of body size with concentrations of Cu and Ag in the bivalve *Macoma balthica*. *Can. J. Fish. Aquat. Sci.* **38**: 1059-1064.
- ¹⁰¹Lobel, P.B., D.A. Wright, 1982a. Relationship between body zinc concentration and allometric growth measurements in the mussel *Mytilus edulis*. *Mar. Biol.* **66**: 145-150.
- ¹⁰²Lobel, P.B., D.A. Wright, 1982b. Gonadal and nongonadal zinc concentrations in mussels. *Mar. Pollut. Bull.* **13**: 320-323.
- ¹⁰³Lowe, D.M. and M.N. Moore, 1979. The cytochemical distributions of zinc (Zn 11) and iron (Fe 111) in the common mussel, *Mytilus edulis*, and their relationship with lysosomes. *J. mar. biol. Ass. U. K.* **59**: 851-858.
- ¹⁰⁴Cossa, D., E. Bourget and J. Piuze, 1979. Sexual maturation as a source of variation in the relationship between cadmium concentration and body weight of *Mytilus edulis* L. *Mar. Pollut. Bull.* **10**: 174-176.
- ¹⁰⁵Watling, H.R., R.J. Watling, 1976. Trace metals in *Chorornytilus mendionalis*. *Mar. Pollut. Bull.* **7**: 91-94.

Eustrongylides sp. (Nematoda: Dioctophymatoidea) Occurrence in the Fourspine Stickleback, *Apeltes quadracus*

by

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Larvae of *Eustrongylides* have been reported in 14 orders of fish worldwide, including Gasterosteiformes¹, but not in the fourspine stickleback^{2,3}. This fish occurs from Labrador⁴ to Virginia⁵. It is abundant in salt water, but sometimes enters fresh waters⁴. A collection of 143 stickleback (114 males, 29 females) were collected from a 575 meter X 75 meter fresh water pond with a mud bottom (Fig. 1) located on Sandy Hook, New Jersey. The pond is surrounded by Phragmites (Fig. 2). Fish were collected both with a 20 foot X six foot 3/8 inch mesh seine and with long-handled dip nets with 1/4 inch mesh on several occasions in September and October, 2006. Fish averaging 38 mm total length (range 32-52 mm) were sacrificed and their viscera removed and examined (Fig. 3). Five males (4%) and one female (3%) contained at least one nematode (Fig. 4). One fish contained two nematodes. The average length of the worms was 40 mm. One other fish species, *Fundulus heteroclitus*, the mummichog, occurs in the pond. Examination of the viscera of 44 mummichog (25 males, 19 females) yielded a nematode infection rate of 56% for males and 47% for females. Infected fish of this species contained one to 15 worms (the average length of the worms was 90 mm). The pond is utilized by several bird species that are hosts for adult *Eustrongylides* sp. The nematode eggs, passed by bird feces deposited in the pond, are eaten and develop in tubificed oligochaetes, the first intermediate hosts⁶. The infected oligochaetes are then eaten by some of the fishes in the pond, the second intermediate hosts of *Eustrongylides* sp.

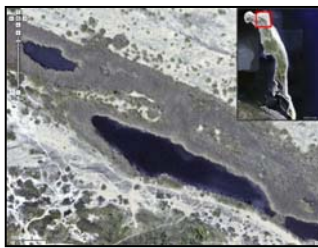


Figure 1. Northpond, a freshwater pond located at Sandy Hook, New Jersey. Insert in upper right, is an aerial view of Sandy Hook. The area outlined in red shows the location of Northpond.



Figure 2. Collecting site at Northpond.

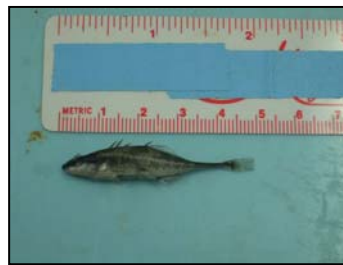


Figure 3. Fourspine Stickleback (female) collected from Northpond.



Figure 4. Nematodes, *Eustrongylides* sp. removed from the Fourspine Stickleback.

References

- ¹Karmanova, E.M., 1968. Dioctophymidea of animals and man and diseases caused by them. *In* Fundamentals of nematology. Vol. 20. Academy of Sciences of the USSR, Moscow. (Translated and published for the U.S. Department of Agriculture by Amerind Publishing Co., Pvt. Ltd., New Delhi, 1985.)
- ²Hoffman, G.L., 1999. Parasites of North American Freshwater Fishes, 2nd Ed. Cornell University Press, New York, NY.
- ³Anderson, R.C., 2000. Nematode Parasites of Vertebrates. Their Development and Transmission. 2nd Ed. CABI Publishing. New York, NY.
- ⁴Blair, W.F., A.P. Blair, P. Brodcorp, F.R. Cagle, and G.A. Moore, 1957. Vertebrates of the United States. McGraw-Hill, New York, NY.
- ⁵Eddy, S., and J.C. Underhill, 1978. How to know the Freshwater Fishes, 3rd Ed. Wm. C. Brown/McGraw-Hill, New York, NY.
- ⁶Measures, L.N., 1988. The development and pathogenesis of *Eustrongylides tubifex* (Nematoda: Dioctophymatoidea) in fish. *Journal of Parasitology* **74**:294-304.

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The **Treasurer** of the Association is responsible for the preparation of an annual fiscal report, processing of dues, preparing regular financial reports for the Executive Board meetings, income tax reports, and other duties usually pertaining to this office.

The **Members-at Large** shall chair committees (Articulation, Exhibition, etc.) and handle other assignments as directed by the Executive Board.

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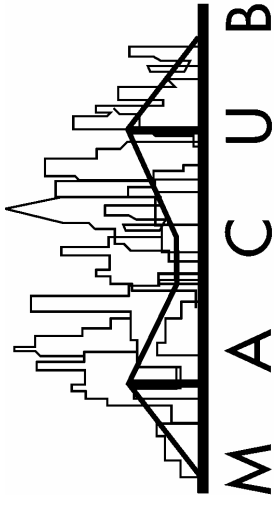
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